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S. A. Chambers

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Birds as Environmental Indicators

Review of Literature

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EXECUTIVE SUMMARY

Under the Signs of Healthy Parks initiative, Parks Victoria seeks to establish a monitoring program throughout the Victorian reserves network to provide information on natural values, threatening processes, emerging conservation issues and the efficacy of management activities. One of the most obvious and readily observable of faunal groups is birds. This report reviews the potential for birds to function as indicator species for other parameters of the environment and assesses the potential for a reliable surrogate monitoring program based on birds to be implemented within the Victorian reserves network.

The report finds that the use of indicator species is contentious and the approach has several limitations. However, indicator species may provide a crude but useful index of environmental parameters when other monitoring is not feasible, provided that precautions are taken in the selection of species and the interpretation of monitoring data. Birds are often claimed to be good indicator species, but in reality they possess qualities that can both enhance and diminish their value as indicator species. In at least some circumstances, birds are likely to be a less suitable target for surrogate monitoring than other taxa.

The report indentifies that monitoring breeding success and/or causes of breeding failure, can provide a more accurate surrogate assessment of environmental properties than presence/absence or abundance data alone, especially in relation to coastal and highly mobile species. However, this typically requires targeted studies to be undertaken where the costs or resources needed for the study could approach or even exceed the costs or resources required to undertake direct sampling of the variable of interest.

Currently, systematic monitoring is conducted in only a small proportion of Victorian reserves and many reserves lack even basic inventories of fauna and flora. In this context, population trends in selected species of birds may alert Parks Victoria to emerging conservation issues in reserves for which other forms of monitoring are not feasible. A set of candidate avian indicator species is identified that may fulfil such a role in light of the identified limitations of the indicator species approach.

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1 INTRODUCTION

1.1. Indicators and Indicator Species

To preserve and manage the natural environment, it is necessary to monitor the condition of many constituent ecosystems. Ideally, conservation agencies would like to monitor all properties of interest within a natural ecosystem. In reality, finite resources and the vast complexity of natural ecosystems permit only a small proportion of these properties to be monitored. In the absence of comprehensive data on whole ecosystems, conservation agencies may choose to monitor a set of indicators: physical, chemical or biological properties that indicate the status of un-sample parameters of the environment. Often, individual species or groups of species are selected to function as biological indicators.

This report has two aims:

- To review the ability of birds to function as indicator species in light of the published scientific literature.
- To determine the prospects for establishing a reliable surrogate monitoring program based on birds within the Victorian reserves network.

In the context of this report, the following definition of indicator species has been adopted:

"A species whose characteristics (e.g. presence or absence, abundance, density, mortality rate, breeding success) indicate the condition of ecosystems, the status of other taxa, the presence and impacts of stressors, or patterns of biological diversity" (after Landres et al. 1988 and McGeoch 1998).

1.2 A Note on Definitions

The term indicator is routinely applied in the scientific literature to both a species to be monitored (*i.e.* the indicator species) and the specific characteristic of the indicator species that informs the status of the ecological property of interest (*e.g.* presence or absence, density, mortality rate, reproductive success). For example, Carignan & Villard (2002) state "...the Ovenbird, *Seiurus aurocapillus*, is a good indicator of closed-canopy, mature forests with a sparse understorey", when in fact "...the [presence/absence of the] Ovenbird...is a good indicator of [the presence/absence of] closed-canopy, mature forests with a sparse understory". This report uses the term 'indicator species' where applicable to avoid confusion.

2 INDICATOR SPECIES

2.1 Types of Indicator Species

Indicator species can be classified in many different ways (Lindenmayer & Burgman 2005) according to their function. For example, McGeoch (1998) recognised three broad classes of indicator species:

- Environmental indicators indicate the presence and/or intensity of stressors. McGeoch (1998) included in this class five types of pollution indicator species identified by Spellerberg (1991):
 - 1. Sentinels are sensitive species that are introduced into a target environment, for example, to provide early-warning of the presence of pollutants or to determine the effects of pollutants on biota.
 - 2. Detectors are species that are endemic to a target area and may exhibit a measurable response to change in their environment, *e.g.* changes in behaviour, mortality or age-class structure.
 - **3. Exploiters** are species whose presence indicates probable disturbance or pollution. Exploiter species often thrive in disturbed or polluted areas because competitors are unable to persist in the altered environment and/or because they can utilize polluted ecosystems or the pollutants themselves.
 - **4.** Accumulators are species that accumulate pollutants in measurable quantities in their body tissue.
 - **5. Bioassay organisms** are used in laboratory studies to detect the presence and/or concentration of pollutants or to determine pollutant toxicity.
- Ecological indicators indicate the impacts of stressors on other taxa.
- Biodiversity indicators indicate the diversity of other taxa within a habitat or set of habitats.

However, Lindenmayer & Burgman (2005) recognised five main classes of indicator species (after Lindenmayer *et al.* 2002):

- Bio-indicator species respond to changes in the environment. For example, the North American species Kirtland's Warbler *Dendroica kirtlandii* has been proposed as an indicator of global warming (Botkin *et al.* 1991).
- Site-type indicator species indicate the presence of particular environmental conditions, such as certain rock or soil types or habitat elements at a particular site.
- Recovery indicator species reflect the extent of recovery of an ecosystem (Lindmenmayer & Burgman 2005). For example, birds have been used to determine the success of habitat restoration programs in the United States (Weller 1995; Gardali *et al.* 2006).
- **Management indicator species** indicate the impacts of management regimes on biota. For example, the return of palatable plant species after the removal of grazing mammals.
- Pollution indicator species indicate the effects of pollutants on the biotic environment. Lindenmayer & Burgman (2005) included in this class the five types of pollution indicator species identified by Spellerberg (1991) and described above.

Lindenmayer & Burgman (2005) also acknowledged keystone species and dominant species in their discussion of the indicator species concept. The defining characteristics of these species make them attractive targets for environmental monitoring, but they need not necessarily function as indicator species as defined in this report:

- Keystone species interact strongly with sympatric taxa. Because of these interactions, keystone species may have a disproportionately large influence on natural ecosystems relative to their abundance (Mills *et al.* 1993; Paine 1995; Simberloff 1998). For example, the Southern Cassowary *Casuarius casuarius* is considered a keystone species because it is the main dispersal agent for approximately 100 fruiting plant species in rainforests of north-eastern Queensland (Moore 2007).
- Dominant species comprise a large proportion of the biomass of an ecosystem or numerically dominate an ecosystem (Lindenmayer *et al.* 2002; Lindenmayer & Burgman 2005).

The indicator species concept is often also associated with other taxon-based surrogate approaches in the literature. Like keystone and dominant species, the surrogates listed below possess attributes that are attractive for environmental monitoring, but they may not necessarily function as effective indicator species:

- Umbrella species require large areas of habitat to maintain viable populations. By conserving habitat for umbrella species, it is presumed that sympatric species with less demanding spatial habitat requirements will also be protected (Simberloff 1998; Carignan & Villard 2002; Favreau *et al.* 2006).
- Focal species are the subject of a taxon-based surrogate conservation approach developed by Lambeck (1997). The focal species approach builds on the concept of umbrella species by (1) identifying the threatening processes that are operating in an ecosystem or landscape; (2) selecting a suite of focal species, each of which is considered to be the most sensitive species in the ecosystem or landscape to one or more of the identified threats; and (3) using the requirements of the nominated focal species to set conservation targets (Lambeck 1997; Favreau *et al.* 2006). Lambeck (1997) recognised four sub-classes of focal species:
 - **1. Area-limited 'umbrella' species** require large areas of habitat to maintain viable populations (Carignan & Villard 2002).
 - 2. Dispersal-limited species are limited in their ability to move from one patch of habitat to another or face a high risk of mortality in doing so (Carignan & Villard 2002).
 - **3.** Resource-limited species depend on specific resources (*e.g.* nectar, fruit, tree hollows) that may be spatially or temporally scarce (Carignan & Villard 2002).
 - Process-limited species are sensitive to ecological processes such as fire, flood, grazing, predation or competition with invasive species (Lambeck 1997; Carignan & Villard 2002).
 - 5. Flagship species are charismatic species that easily attract public support for conservation (*e.g.* Giant Panda *Ailuropoda melanoleuca* and whales; Carignan & Villard 2002; Favreau *et al.* 2006).

2.2 Characteristics of Indicator Species

To be considered for use in an environmental monitoring program, an indicator species must possess two essential qualities:

- The indicator species must be sensitive to change in the environmental property or properties of interest (Landres *et al.* 1988; Simberloff 1998). This association is intuitively necessary for the status of the environmental property or properties of interest to be inferred from trends in the indicator species.
- The indicator species must be sufficiently detectable and the targeted characteristic(s) of the indicator species sufficiently mensurable for monitoring data to be collected in a reliable and repeatable fashion. This attribute is necessary to ensure that sample sizes are large enough for change to be detected.

An indicator species may also possess a combination of non-essential but desirable attributes that enhance their utility. Examples of these qualities are provided in Table 1 below.

Indicator Species Attribute	Importance for monitoring	References
Resolved Taxonomy	This is stable and well-resolved which minimises the potential for monitoring data to be confounded by misidentification of species in the field	Pearson & Cassola 1992; Furness <i>et al.</i> 1993; Carignan & Villard 2002; Gregory <i>et al.</i> 2005
Well understood Biology, ecology and life history	A comprehensive understanding of these elements is necessary to correctly interpret monitoring data	Landres <i>et al.</i> 1988; Noss 1990; Pearson & Cassola 1992. Furness <i>et al.</i> 1993
History of previous monitoring	Where available, historical monitoring data provides a baseline against which contemporary data can be compared and monitoring methods that are tried and tested.	
Broad distribution	This permits multiple sites to be monitored and compared using a single indicator species although comparisons between sites can be problematic	Noss 1990; Pearson & Cassola 1992. Landres <i>et al.</i> 1988.
Resident species	Recommended because non-resident species such as migrants or nomads (1) may be able to avoid the impacts of stressors operating within the target area by utilising resources elsewhere and (2) may be impacted by stressors operating outside of the target area. An additional bonus of targeting resident species is that direct monitoring can be conducted throughout the year Although resident species are the preferred targets for ecological monitoring, the inclusion of both resident and non- resident species in a monitoring scheme may enable a greater range of stressors to be detected than would be possible by monitoring resident species alone	Szaro & Balda 1982; Bock & Webb 1984; Landres <i>et al.</i> 1988; Koskimies 1989; Hilty & Merenlender 2000. Koskimies 1989.

Table 1. Desirable attributes of an indicator species used for environmental monitoring.

Indicator Species Attribute	Importance for monitoring	References
Detectability	Easy to detect and inexpensive to monitor	Noss 1990; Pearson & Cassola 1992
Specialised species	Species that rely on a limited range of habitat types and/or food resources are likely to be more sensitive to disturbance than generalist species. This is because generalist species may be able to avoid or lessen the impacts of disturbance by switching food resources or altering patterns of habitat use	Koskimies 1989; Pearson & Cassola 1992; Hilty & Merenlender 2000; Carignan & Villard 2002
Tolerance	The tolerance of the indicator species to change in the environmental property or properties of interest has been determined and well understood	Hilty & Merenlender 2000
Species exhibits consistent response to changes and threats	The indicator species exhibits minimal variation in its response to change in the environmental property or properties of interest	Landres <i>et al.</i> 1988
Provides early- warning	The indicator species provides early-warning of change in the environmental property of interest	Kelly & Harwell 1990; Noss 1990
Directly indicates the cause of change	The indicator species directly indicates the cause of change rather than simply the existence of change	Morrison 1986; Carignan & Villard 2002
Species responds to multiple stressors	Indicator species provide a means for continuous assessment of multiple stressors operating over a broad range of intensities. This permits multiple stressors to be monitored using a single indicator species and ensures that the indicator species will not become ineffective beyond certain stressor intensity thresholds	Noss 1990; Woodley 1996; O'Connell 1998; Gibbs <i>et al.</i> 1999
Social, political and economic importance	The indicator species is aligned with relevant social, political or economic agendas	Pearson & Cassola 1992; Hilty & Merenlender 2000
Generation time	Species with short generation times (<i>e.g.</i> invertebrates) tend to react more quickly to ecosystem changes than species with longer generation times. It can be useful when the generation time of the indicator species matches the aims of the monitoring program	Niemelä et al. 1993; Carginan & Villard 2004. Peters 1983.

Table 1.	(continued)
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No single indicator species will possess all of the desired attributes identified (Noss 1990). Indeed, some desired attributes conflict with one another, for example, cosmopolitan distribution versus habitat specialism (Hilty & Merenlender 2000).

An issue that may be encountered when assessing candidate indicator species is that many of the desirable attributes listed above assume a sound and detailed knowledge of the biology and ecology of the species in question. However, this is often not the case. In instances where scientific data are lacking, it may be necessary to judge the merit of candidate indicator species based on alternative sources of information such as expert opinion, personal observations or non-scientific literature, pending additional research.

2.2.1 Control Populations

One additional factor to consider when discussing the desired qualities of an indicator species is the existence of 'control' populations. In the context of monitoring indicator species to determine the presence or impacts of stressors, it may prove instructive to compare populations in target areas in which the stressor or stressors of interest are known or suspected to be operating ('treatment' populations) with populations in similar habitat in non-target areas in which the stressor or stressors of interest are known to be absent ('control' populations). However, caution must be exercised if such comparisons are made. As Landres *et al.* (1988) note, spatially-separated ecosystems that appear similar may actually differ in vegetation structure, floral and/or faunal species composition, spatial patterning of habitats and resources or natural disturbance regimes. Any of these elements may influence the status or role of an indicator species in an ecosystem.

2.3 Selection of Indicator Species

Many procedures have been proposed to select indicator species (Carignan & Villard 2002). For example, Hilty & Merenlender (2000) proposed a step-wise filtering process based on 13 selection criteria recommended by nine published studies. Hutto (1998) identified candidate indicator species based on their occurrence within one or a few *a priori*-defined habitat types. Kremen (1992), Dufrêne & Legendre (1997) and Hausner *et al.* (2003) used statistical techniques to test for associations between species and particular habitat attributes.

Carignan & Villard (2002) considered that indicator species for monitoring environmental condition should be selected by quantitative criteria in preference to more subjective qualitative criteria. They proposed a simple procedure with two quantitative selection criteria: (1) the frequency of occurrence of the indicator species differs among areas with contrasting degrees of anthropogenic disturbance; and (2) the indicator species is a habitat specialist.

2.4 Benefits of Using Multiple Indicator Species

Many contemporary authors advocate the use of multiple indicator species or groups of indicator species (Koskimies 1989; Noss 1990; Kremen 1994; Griffith 1997; Hutto 1998; Carignan & Villard 2002; Kavanagh *et al.* 2005). This is because one or a small number of indicator species is unlikely to adequately represent the full range of biotic responses to a particular set of environmental conditions (Carignan & Villard 2002). Furthermore, multiple species can usually be monitored by common methods, providing additional data at little or no extra cost (Hutto 1998; Kavanagh *et al.* 2005).

3 BIRDS AS INDICATOR SPECIES

3.1 Use of Birds as Indicator Species

Birds have been proposed, assessed or used as indicator species for a range of environmental parameters, including the following:

- Biodiversity and species richness (Block *et al.* 1987; Jannson 1998; Blair 1999; Chase *et al.* 2000; Vielliard 2000; Mikusinski *et al.* 2001; Juutinen & Mönkkönen 2004; Kati *et al.* 2004; Sauberer *et al.* 2004; Fleishman *et al.* 2005; Mattsson & Cooper 2006; Sergio *et al.* 2006), including patterns of occurrence of rare and threatened species (Balmford & Garson *et al.* 2002; Lawler *et al.* 2003; Thomson *et al.* 2007). In some instances, birds have been used as surrogate taxa for biodiversity to prioritise areas for conservation (Baltzer *et al.* 2000; Moore *et al.* 2003; Jiguet & Julliard 2006; Loyola *et al.* 2007).
- Environmental contamination by pollutants such as pesticides (especially persistent organochlorines), heavy metals and polychlorinated biphenyls (Brisbin 1993; Furness 1993; Wren *et al.* 1994; Dauwe *et al.* 2002; Mochizuki *et al.* 2002; Boncompagni *et al.* 2003; Matz & Parsons 2004; DeWitt *et al.* 2006; Papp *et al.* 2007).
- The condition of ecosystems (Kushlan 1993; O'Connell 2000; Tankersley 2004; Zockler 2005) including forests (Canterbury *et al.* 2000; Uliczka & Angelstam 2000; Becker & Agreda 2005; Venier & Pearce 2005); rainforests (Lee *et al.* 2005); grasslands (Browder *et al.* 2002); rangelands (Bradford *et al.* 1998; Whitford *et al.* 1998); riparian ecosystems (Croonquist & Brooks 1991; Bryce *et al.* 2002); terrestrial wetlands (Tamisier & Boudouresque 1994; Sorace *et al.* 1999, 2002; O'Connor *et al.* 2000; Paillisson *et al.* 2002; Deluca *et al.* 2004); marine ecosystems (Powell & Powell 1986; Furness & Nettleship 1991; Mallory *et al.* 2006; Sydeman *et al.* 2007); and urban areas or mosaics (Jedicke 2000; Reynaud & Thioulouse 2000).
- Ecosystem responses to disturbances and processes including urban expansion (Lee *et al.* 2005), logging regimes (Holmes *et al.* 2004; Venier & Pearce 2004; Kavanagh *et al.* 2005), hydrological regimes (Paillisson *et al.* 2002; Desgranges *et al.* 2006), eutrophication (Fernandez *et al.* 2005), replacement of endemic ecosystems with plantations (Hausner *et al.* 2003), grazing (Bock & Webb 1984; Sedgwick & Knopf 1987), hunting (Paillisson *et al.* 2002) and habitat restoration programs (Weller 1995; Gardali *et al.* 2006).

Birds have been utilised as indicator species by government agencies in Australia and elsewhere. The best examples perhaps come from Europe and North America.

In the United Kingdom, composite bird indices are one of 20 'framework' indicators (and one of 68 indicators in total) used to measure progress towards a government goal of achieving sustainable development by 2020. Composite indices for farmland birds, woodland birds, coastal/estuarine birds and wintering waterbirds are produced in a collaborative effort by the British Trust for Ornithology (BTO), The Royal Society for the Protection of Birds (RSBP) and the Department for Environment, Food and Rural Affairs (DEFRA) and are mainly based on abundance data obtained via the Common Bird Census, Breeding Bird Survey and Wetland Bird Survey (H. M. Government 2005; http://www.bto.org/research/indicators/uk_indicators.htm).

The Pan-European Common Bird Monitoring Scheme (PECBMS) uses composite indices for common birds to track the state of biodiversity across Europe. In 2007, PECBMS common bird indices were updated with abundance data on 124 species of birds obtained from monitoring programs in 20 countries (PECBMS 2007).

In North America, the United States Department of Agriculture Forest Service are required by internal policy to identify and monitor 'management indicator species' within each national forest. The management indicator species are monitored to determine the impacts of

management regimes on forest communities (U.S. Congress Office of Technology Assessment 1992).

There are few documented examples of the use of birds as indicator species by government agencies in Australia. At national and state/territory scales, birds have been utilised as indicator species for State of the Environment (SoE) reporting (http://www.environment.gov.au/soe/index.html). For example, 'seabird populations' and the 'abundance and distribution of waterbirds' were two of 263 indicators used to assess the state of the Australian environment for the 'Australia State of the Environment 2006' report (Beeton et al. 2006; http://www.environment.gov.au/soe/index.html) and the 'current distribution of threatened terrestrial bird species' was one of ten indicators used to assess the state of biodiversity in Western Australia for the 'Environment Western Australia 1998: State of the Environment Report' (Government of Western Australia 1998).

At the regional scale, at least one government project in Victoria has used birds as indicators. The Wimmera Catchment Management Authority in western Victoria is currently coordinating a five-year project – the Wimmera Bird Monitoring Project (WBMP) – that aims to use bird monitoring as a means to monitor regional biodiversity and the condition of catchment ecosystems. However the WBMP is not targeting selected indicator species; instead, the project aims to sample region-wide avian diversity, although analysis is likely to focus particularly on two woodland species, Jacky Winter *Microeca fascinans* and Hooded Robin *Melanodryasi cucullata*, that are (1) in regional decline, (2) sensitive to habitat change and (3) considered to be good indicators of woodland condition (Birds Australia 2003; Jon Starks, *pers. comm.*; http://wcma.vic.gov.au).

Although there has been only limited actual use of birds as indicator species by government agencies in Australia, a variety of Australian birds have been proposed as potential indicator species in the literature because of their association with certain conditions and/or sensitivity to certain stressors (Appendix 1). Furthermore, a number of focal species studies have been conducted in Australia, most of which have focused on birds (Huggett 2007). The focal species identified by these studies could be regarded as potential candidates for monitoring because of their demonstrated sensitivity to certain threatening processes (examples of avian focal species identified by studies conducted in southern Australia are provided in Appendix 1). However, the use of focal species as indicator species could prove problematic.

An intrinsic problem of the focal species approach is the difficulty of determining which species is most sensitive to a particular threat (Freudenberger *et al.* 2001; Lambeck 2002; Lindenmayer *et al.* 2002). For example, rare species that are excluded from focal analyses because of a lack of data may be more sensitive to a particular threat than the nominated focal species (Short & Parsons 2004). Furthermore, birds are likely to be less suitable as focal species for fragmentation and isolation of habitat than less mobile taxa such as plants, amphibians and reptiles (Freudenberger *et al.* 2004). Consequently, nominated focal species such as those listed in Appendix 1 may be capable of indicating the presence and impacts of threatening processes, but it is possible that other species or taxa excluded from or rejected during focal analyses may be more sensitive to the threatening process in question. In the context of the indicator species concept, this means that a species overlooked or rejected during focal analyses could be affected by a threatening process, but the threatening process would not be detected or redressed if the process was operating at an intensity below the threshold sensitivity of the focal species.

It is worth noting that work has commenced to develop composite indices to determine population trends of birds in Australia. Cunningham & Olsen (2007) have designed a set of composite statistical measures that make use of volunteer-collected presence-absence data from the Atlas of Australian Birds. Some preliminary trials have taken place in temperate woodlands to identify a suite of representative species to monitor, but substantial work is still needed before a full set of suitable indicator species can be selected.

3.2 Advantages and Disadvantages of Using Birds as Indicator Species

Birds are often considered to be good indicators of the general condition of the environment, although doubt exists over their ability to directly and rapidly indicate changes in ecosystem properties and the impacts of such changes on other taxa (Morrison 1986; Temple & Wiens 1989; Mac Nally *et al.* 2004; Gregory *et al.* 2005). There are both positive and negative aspects to using birds as indicator species. The following attributes are considered to enhance the value of birds as indicator species:

- Birds are easy to detect and observe (e.g. birds are often the most conspicuous faunal taxon in an ecosystem; many species are diurnal and/or brightly-coloured and/or advertise their presence by call) (Hutto 1998; Carignan & Villard 2002; Mac Nally *et al.* 2004).
- The taxonomy of birds is well-resolved and species are generally easy to identify in the field (Furness *et al.* 1993; Gregory *et al.* 2005).
- Birds are widely distributed and occupy a broad range of habitat types and ecological niches.
- The distribution, biology, ecology and life history of birds are well known compared with other taxa (Furness *et al.* 1993; Gregory *et al.* 2005). For example, since 1998 the Atlas of Australian Birds has collected more than 7,300,000 bird records obtained from more than 433,000 surveys across Australia.
- Birds are typically positioned at or near the top of the food chain. This makes birds sensitive to changes at lower levels of the food chain and to environmental contaminants (*e.g.* persistent organochlorines) that accumulate at each level of the food chain (Furness *et al.* 1993; Mac Nally *et al.* 2004; Gregory *et al.* 2005).
- Many bird species pollinate or disperse the seeds of plants. These species may be directly or indirectly linked to the fitness of many other species and play a critical role in the maintenance of natural ecosystems.
- Birds are of interest and concern to both the public and decision-makers (Mac Nally *et al.* 2004; Gregory *et al.* 2005). This interest generates strong support for conservation programs that involve birds and provides an opportunity for skilled volunteers to be recruited from a large number of amateur birdwatchers. The inclusion of volunteer personnel in a monitoring program may reduce costs or facilitate an increase in the scope of the monitoring program (Furness *et al.* 1993).
- Survey techniques for birds are comparatively simple and are capable of capturing information on a multitude of species simultaneously (Hutto 1998).
- Birds are generally less expensive to monitor than other taxa such as invertebrates, reptiles and mammals (Landsberg *et al.* 1999; Mac Nally *et al.* 2004).

The following attributes can either be beneficial or detrimental depending on the spatial and temporal scales and aims of monitoring:

Birds are highly mobile. This allows monitoring to be conducted over broad spatial scales but may make it difficult to link responses of birds to specific conditions or stressors on the ground (Furness *et al.* 1993; Mac Nally *et al.* 2004; Gregory *et al.* 2005). Furthermore, because birds are exceedingly more mobile than almost all other terrestrial taxa, and because many species of bird make greater use of resources in three-dimensional space than other animals, the responses of birds may not accurately reflect the responses of more sedentary, spatially-restricted faunal taxa such as invertebrates, reptiles and small mammals (Mac Nally *et al.* 2004; Gregory *et al.* 2005).

Generation times of birds range from a few years to a few decades in length (Garnett & Crowley 2000). This means that birds are capable of signalling impacts over a period of time (*i.e.* the effects of long-term exposure), but are probably less suitable for indicating short-term disturbances than other taxa with shorter generation times (Furness *et al.* 1993; Gregory *et al.* 2005).

The following attributes are considered to diminish the value of birds as indicator species:

- Birds often respond to secondary or tertiary effects of stressors (Morrison 1986; Koskimies 1989; Temple & Wiens 1989). For example, the El Niño-Southern Oscillation (ENSO) phenomenon has a severe but indirect impact on some species of fish-eating birds in the Pacific Ocean. During ENSO events, a current of cold, nutrient-rich waters originating from southern latitudes (the Humboldt Current) is replaced by a current of warm, nutrient-poor waters originating from the tropics. This change has no direct effect on fish-eating birds, but the nutrient-poor waters associated with ENSO events sustain substantially fewer fish than the nutrient-rich waters of the Humboldt Current. This reduction in the availability of food may lower breeding success and cause crashes in some populations of fish-easting birds. (Barber & Chavez 1983; Schreiber & Schreiber 1984; Duffy 1990). In situations such as this, time lags exist between the onset of the stressor and the response of birds. The existence of time lags is undesirable because (1) the impact of the stressor may be more difficult to mitigate or irreversible by the time a problem is detected and (2) time lags may make it more difficult to trace the correct cause of an indicator species' response (Temple & Wiens 1989).
- In certain circumstances, some opportunist birds may respond positively to environment degradation (Gregory *et al.* 2005), at least initially. For example, the widespread clearing of eucalypt forests and woodlands in eastern Australia since settlement has allowed Noisy Miners to colonise areas that were formerly unsuitable for this species (Catterall *et al.* 2002; Clarke *et al.* 2006).
- Birds possess behavioural and physiological traits that may make them less sensitive to ecosystem changes than some other taxa. For example, birds can regulate fat stores and metal concentrations in their body tissues to a much greater degree than invertebrates. Traits such as these help to buffer birds against the impacts of ecosystem changes and, consequently, may limit the ability of birds to indicate ecosystem changes and their effects on other taxa (Furness *et al.* 1993).

4 BENEFITS AND DEFICIENCIES OF THE INDICATOR SPECIES APPROACH

A primary motivation for using indicator species as a tool to assess the condition of ecosystems without looking at all elements, is an assumption of their ability to provide information on properties of the environment that are otherwise difficult, inconvenient or expensive to be measured directly (Landres *et al.* 1988). For this reason, the indicator species approach holds tremendous appeal for conservationists, land managers and politicians (Carignan & Villard 2002).

However, there are four critical problems that undermine the validity of the indicator species concept:

- Species can respond in different ways to the same environmental conditions or stressors (e.g. Woodwell & Rebuck 1967; Mannan et al. 1984; Block et al. 1987; Reader 1988; Thiollay 1992; Niemelä et al. 1993; Taper et al. 1995; Dormann et al. 2007; Ficetola et al. 2007). This is not surprising considering that all species vary to some degree in their behaviour, habitat requirements and/or life history traits (Mannan et al. 1984; Verner 1984; Szaro 1986; Block et al. 1986; Landres et al. 1988; Martin & Li 1992; Martin 1995; Lindenmayer 1997). In effect, each individual species is exposed to a unique set of environmental conditions and ecological interactions (Landres et al. 1988; Koskimies 1989). This variation in strategies and responses suggests that it is unlikely that any one species will act as a perfect indicator for any other species (Landres et al. 1988; Koskimies 1989; Taper et al. 1995; Carignan & Villard 2002).
- Birds and other animals tend to respond in similar ways to the many different factors that may influence their populations (*e.g.* quality or quantity of habitat, predation, competition, diseases, parasites, weather conditions and natural stochastic fluctuations). This can make it difficult to accurately identify the root cause of an indicator species response (Steele *et al.* 1984; Morrison 1986; Landres *et al.* 1988; Koskimies 1989; Carignan & Villard 2002).
- The impact of disturbance on a species may not become fully apparent until sometime after the onset of the disturbance. For example, a population of Splendid Fairy-wrens in Western Australia began to decline more than three years after a major fire, evidently because of reduced breeding productivity and the replacement of experienced breeders with novices, which were less productive and suffered higher rates of mortality (Russell & Rowley 1993). The potential for species to exhibit delayed responses to disturbance means it may be possible for a disturbance to go undetected by indicator species monitoring until its impacts are difficult to reverse or, worse, are irreversible (Temple & Wiens 1989; Lindenmayer *et al.* 2000).
- To indicate biodiversity or the status of other taxa, indicator species must be highly congruent with the taxa for which they are proposed to be indicative (Lindenmayer 1999; Kavanagh & Stanton 2005). However, many studies have collectively demonstrated that species-rich areas for one taxon do not consistently coincide with species-rich areas for other taxa at sub-global scales (*e.g.* Prendergast *et al.* 1993; Prendergast & Eversham 1997; Oliver *et al.* 1998; Mikusinski *et al.* 2001; Lund & Rahbek 2002; Söderström *et al.* 2002; Vessby *et al.* 2002; Kati *et al.* 2004; Sauberer *et al.* 2004; Schulze *et al.* 2004; Grenyer *et al.* 2006; Similä *et al.* 2006; Pawar *et al.* 2007; Pearman & Weber 2007).

5 CONCLUSIONS AND FUTURE DIRECTIONS

The concept of using a set of indicator species to monitor the condition of the environment possesses undoubted appeal for conservationists, land managers and politicians alike. However, in reality, the ability of birds and other taxa as useful surrogates for other properties of the environment is debatable (Landres *et al.* 1988; Carignan & Villard 2002; Lindenmayer & Burgman 2005).

Direct monitoring of environmental properties of interest should be undertaken wherever possible (Landres *et al.* 1988), but ultimately, the finite resources available to conservation agencies and the vast complexity of natural ecosystems make the use of indicators virtually unavoidable (Simberloff 1998). If indicator species are to be used, it is important to recognise their limitations. As Hutto (1998) stated, there is little reason to expect that a small group of indicator species will provide much more than a crude index of ecosystem condition.

Despite problems associated with their use, Carignan & Villard (2002) considered that indicator species can be a useful tool for conservation if (1) many species representing various taxa and life histories are included in a monitoring program, (2) their selection is primarily based on sound quantitative data from the region of interest and (3) caution is exercised when interpreting indicator trends to separate actual signals from fluctuations that may be unrelated to ecosystem deterioration.

At present, very little systematic monitoring of natural attributes is undertaken within the Victorian reserves network. Indeed, many reserves lack basic inventories of fauna and flora. In this context, the establishment of a state-wide monitoring program for birds would provide valuable information on avifaunal populations and, in instances where no other data exists, might also provide insights into ecosystem character (*e.g.* habitat types necessary to sustain recorded bird species). The observations of ecosystem characteristics can be further utilised to provide assessment information on land management issues, such as degradation of habitat from over grazing; or perhaps overcrowding of the understorey from weed invasion, which can assist with identifying appropriate management strategies.

For reserves in which direct monitoring of environmental parameters other than avifauna is not feasible, it is possible that trends in selected species of birds could alert Parks Victoria to conservation issues that may otherwise go undetected, although monitoring of these species or groups can not be guaranteed to capture all threatening processes operating within a reserve, or to reflect impacts on other taxa. It is important to stress that targeting selected species or groups of birds alone is a non-ideal option for monitoring reserve values, but may be useful when the alternative is no monitoring at all.

In the following section, a set of candidate avian indicator species is nominated that may provide some indication of reserve values in the absence of other forms of monitoring.

6 CANDIDATE INDICATOR SPECIES

A set of potential indicator species for Victorian reserves is presented below and in Table 2. Species were identified based on expert opinion, peer-reviewed literature and/or grey literature. Species were selected because they were considered to be associated with one or more environmental elements or processes and because they possessed some combination of the desirable attributes listed in Table 2. Generally, it was not possible to assess candidate species against the other desirable attributes described in Section 3 of this report.

Candidate Species	Easy to Detect	Well- studied	Resident	Specialist	Prior Monitoring
Australian Owlet-nightjar			+		
White-browed Treecreeper	+	+	+	+	
Brown Treecreeper	+	+	+		
Speckled Warbler			+	+	
Noisy Miner	+	+	+		
Hooded Robin	+	+	+		
Diamond Firetail			+	+	
Malleefowl		+	+	+	+
Mallee Emu-wren			+	+	+
Yellow-throated Miner	+	+	+		
Chestnut Quail-thrush			+	+	
Red-kneed Dotterel	+			+	+
Red-capped Plover	+	+	+		+
Ground Parrot		+	+	+	+
Southern Emu-wren		+	+	+	
Banded Lapwing	+				
Sooty Owl		+	+	+	+
Superb Lyrebird	+	+	+		+
Red-browed Treecreeper	+		+		
Rose Robin					
Olive Whistler	+		+		
Azure Kingfisher			+	+	
Striated Fieldwren	+		+		
Eastern Curlew	+	+			+
Australian Pied Oystercatcher	+	+	+		+
Hooded Plover	+	+	+	+	+

Table 2. Beneficial attributes of candidate indicator species for the Victorian reserves system.

The 'properties indicated' component of the species profiles presented below lists the environmental properties that the candidate species are known or believed to be associated with. The 'justification' component of the species profiles summarises the evidence for associations between the candidate species and the environmental properties they are proposed to indicate.

In most instances, rigorous dedicated studies are needed to validate relationships between candidate indicator species and the environmental properties they are proposed to indicate (Lindenmayer 1999). Some of the candidate indicator species listed below may only prove useful if monitoring target parameters other than the presence/absence of the species (*e.g.* rates of breeding success, causes of breeding failure). These parameters would typically require targeted studies to be measured accurately. It is highly plausible that in some situations the costs or resources needed to complete targeted studies of an indicator species could approach or be equal to or exceed the costs or resources required to undertake direct sampling of the variable of interest. In such circumstances, managers should undertake direct sampling of the variable of interest.

Australian Owlet-nightjar Aegotheles cristatus

Natural Ecosystem Group: Dry Forest and Woodlands, Mallee.

Properties Indicated: Hollow-bearing (*i.e.* mature) trees.

Justification: Australian Owlet-nightjars usually nest and roost in small hollows in trees and tree-stumps (Hollands 1991; Brigham & Geiser 1997; Higgins 1999). However, they sometimes also nest and/or roost in cavities in other substrates including river-banks, cliff-faces and man-made structures (*e.g.* nest boxes, fence posts, pipes and crevices in buildings; see summary in Higgins 1999), which suggests that the presence of Australian Owlet-nightjars may not perfectly indicate the presence of hollow-bearing trees in all instances. A wide range of other fauna species require small hollows for nesting and shelter and in some habitats, such as in mallee woodlands, the presence of hollow-dependent species can indicate mature habitat.

White-browed Treecreeper Climacteris affinis

Natural Ecosystem Group: Dry Forest and Woodlands (semi-arid).

Properties Indicated: Hollow-bearing (*i.e.* mature) trees and coarse woody debris; habitat connectivity.

Justification: In Victoria, White-browed Treecreepers selectively inhabit open, semi-arid woodlands dominated by Belah *Casuarina cristata* or Slender Cypress-Pine *Callitris gracilis* and Buloke *Allocasuarina luehmannii* (Emison *et al.* 1987; Radford & Bennett 2004). They nest in hollows in dead or partially-dead trees (especially Belah) or occasionally in decaying tree-stumps (Emison *et al.* 1987; Higgins *et al.* 2001). Their foraging sites include coarse woody debris (*i.e.* fallen trees and branches) and tree stumps (Emison *et al.* 1987; Higgins *et al.* 2001).

White-browed Treecreepers are sensitive to loss, fragmentation and degradation of their woodland habitats. For example, they were not detected in remnants that contained less than 18.5 ha of Belah woodland in the Millewa district of north-western Victoria (Radford & Bennett 2006), they require landscape-scale habitat connectivity to facilitate movement between remnants (Radford & Bennett 2004, 2006) and they exhibit a preference for ungrazed or lightly-grazed remnants (Emison *et al.* 1987; Radford & Bennett 2006).

Brown Treecreeper Climacterus picumnus

Natural Ecosystem Group: Dry Forest and Woodlands, Mallee.

Properties Indicated: Hollow-bearing (*i.e.* mature) trees; coarse woody debris; habitat connectivity; patch-size; fire regime.

Justification: Brown Treecreepers usually, but not always, nest and roost in hollows in dead or partially-dead trees (Noske 1977, 1982a, b; Higgins *et al.* 2001). They often forage on coarse woody debris (Noske 1979; Antos & Bennett 2006).

The presence and density of Brown Treecreepers appears to be influenced by the presence and volume of coarse woody debris. Experimental manipulation of fallen timber loads in River Red Gum *Eucalyptus camaldulensis* forest in central-northern Victoria, showed that densities of Brown Treecreepers increased rapidly (*i.e.* within six months of manipulation) and substantially at sites where loads of fallen timber were manually raised to \geq 40 Mg/ha (Mac Nally *et al.* 2002). Follow-up surveys conducted in the same forest revealed that densities of Brown Treecreepers remained elevated at sites where fallen timber loads had been increased to \geq 40 Mg/ha three years after experimental manipulation (Mac Nally 2006). Analysis of four years data (1998-2001) from the Atlas of Australian Birds also revealed a positive association between the presence of Brown Treecreepers and the presence of fallen timber (Barrett *et al.* 2002).

Brown Treecreepers are sensitive to fragmentation of their habitat. They appear unable or reluctant to disperse across fragmented landscapes with non-wooded habitat (Walters *et al.* 1999; Cooper 2002; Cooper *et al.* 2002a, b). For example, in central New South Wales, Brown Treecreepers were not detected in remnants located more than 700 m from the next nearest remnant (Cooper *et al.* 2002). In north-eastern New South Wales, Brown Treecreepers moved within and between patches in contiguous habitat, with a maximum recorded movement of 4.5 km, but moved only within patches in fragmented habitat, with a maximum recorded movement of 1.4 km (Cooper & Walters 2002a).

Brown Treecreepers may indicate the minimum size of remnant woodland patches. For example, surveys in north-eastern and central New South Wales failed to detect Brown Treecreepers in remnants of less than 9 ha and 10 ha respectively (Barrett 1995; Cooper *et al.* 2002).

Brown Treecreepers may also indicate the fire history of their habitat. A study in dry sclerophyll forest in south-eastern Queensland found that Brown Treecreepers preferred sites with intermediate fire intervals (2.5-4.0 years) and were more common in sites with short mean fire intervals than long mean fire intervals. This may be associated with the species' preference for open areas with bare ground and short grasses (Smyth *et al.* 2002).

Speckled Warbler Chthonicola sagittata

Natural Ecosystem Group: Dry Forest and Woodlands.

Properties Indicated: Open understorey; sparse ground cover; coarse woody debris; habitat connectivity.

Justification: Speckled Warblers inhabit dry forests and woodlands with an open understorey and sparse or patchy ground cover (*e.g.* Conole 1981; Ford & Bell 1981; Ford *et al.* 1985; Emison *et al.* 1987; Gosper 1992; Er & Tidemann 1996; Tzaros 1996). They mostly forage among grasses and leaf-litter and on patches of bare ground, often around coarse woody debris (Ford *et al.* 1986; Tzaros 1996). Their nests sometimes abut or are placed among coarse woody debris (Tzaros 1996; Gardner 2002a; Higgins & Peter 2002).

Speckled Warblers are sensitive to habitat fragmentation. The species appears to be/become locally extinct in regions where habitat fragments are less than 100 ha in size (Traill & Duncan 2000). Larger habitat fragments are probably required to maintain viable populations. For example, Gardner (2002b) only recorded the species in remnants of 300 or more hectares. Watson *et al.* (2005) found that a 50% probability of occurrence was only attained by remnants of 180-390 ha, depending on the surrounding landscape. A population viability analysis by Gardner & Heinsohn (2007) predicted that in the absence of an Allee effect (*i.e.* slowed growth in small or low density populations), only high-density populations in remnants of more than 300 ha and low-density populations in remnants of more than 700 ha had more than an 80% probability of persisting beyond 100 years.

Noisy Miner Manorina melanocephala

Natural Ecosystem Group: Dry Forest and Woodlands.

Properties Indicated: Degraded and simplified forests and woodlands, fragmentation.

Justification: Noisy Miners are typically absent from intact, high-quality wooded habitat that supports dense understorey vegetation. They infiltrate dry forests and woodlands only where these exist in small, open patches that lack understorey shrubs, or possess only a sparse layer of understorey shrubs, and/or are located adjacent to more open habitat(s) (*e.g.* Morris 1975; Gepp & Fife 1975; Jones 1986; Emison *et al.* 1987; Clarke *et al.* 1995, 2006; Possingham & Possingham 1997; Catterall *et al.* 2002). The presence of Noisy Miners may therefore indicate that habitat is degraded and fragmented.

Hooded Robin Melanodryas cucullata

Natural Ecosystem Group: Dry Forest and Woodlands.

Properties Indicated: Complex woodlands; coarse woody debris; habitat connectivity; patch-size.

Justification: Hooded Robins inhabit large and intact areas of woodland that have a complex understorey and dead or fallen timber (Sullivan 1993; Fitri & Ford 1997, 2003a, b; Higgins & Peter 2002; Maron & Lill 2005; Antos & Bennett 2006; Ford & Thompson 2006; see below). They are highly sensitive to alteration of their preferred woodland habitats. For example, in the northern region of the Australian Capital Territory and adjacent areas of New South Wales, Freudenberger (1999) found that Hooded Robins required structurally complex woodlands (>20% tree canopy cover with >20% cover of shrubs 0.5-4.0 m in height and >40% cover of ground plants and logs or rocks and leaf litter) that were more than 100 ha in area and that were located within about 1 km of other remnant patches of woodland. In the same region, Watson et al. (2001) only recorded Hooded Robins in structurally complex woodland remnants of 100 or more hectares in area. In south-central New South Wales, Freudenberger & Stol (2002) only recorded Hooded Robins in woodland remnants that were more than 20 ha in area, contained a complex of shrubs and fallen timber, and were located within 1-2 km of other woodland remnants. Because of their sensitivity to habitat alteration, Hooded Robins have been nominated as an indicator species for woodland condition (Ford & Thompson 2006) and targeted by focal species approaches (e.g. Watson et al. 2001).

Diamond Firetail Stagonopleura guttata

Natural Ecosystem Group: Dry Forest and Woodlands.

Properties Indicated: Understorey shrubs and regrowth; native ground cover; habitat connectivity; patch-size.

Justification: Diamond Firetails mostly inhabit open forests and woodlands that are dominated by eucalypts (or sometimes by casuarinas or cypress-pines) and support an understorey of shrubs, small trees and regrowth, and a ground-cover of grasses. They also frequently occur in lightly timbered grasslands and farmlands with scattered or remnant patches of trees (Emison *et al.* 1987; Higgins *et al.* 2006).

Diamond Firetails are sensitive to loss, fragmentation and degradation of grassy woodlands (Ford & Thompson 2005, 2006). For example, Diamond Firetails were selected as a focal species for the Goulburn Broken catchment in north-central Victoria because they were only recorded in woodland remnants more than 20 ha in size, which had patches of dense shrubs and a short cover of native plants, and were located within 1 km of other woodland remnants (Robinson & Howell 2003). In the Boorowa River catchment of New South Wales, Freudenberger (2001) only recorded Diamond Firetails in woodland remnants which were five or more hectares in size, which had a moderately complex understorey, and which were located at most 2.7 km from a woodland remnant more than 10 ha in size. Ford & Thompson (2005, 2006) considered the presence/absence or abundance of Diamond Firetails to be a useful indicator of (grassy) woodland condition.

Malleefowl Leipoa ocellata

Natural Ecosystem Group: Mallee.

Properties Indicated: Fire regime; ground-litter; habitat connectivity; predator densities.

Justification: Malleefowl exhibit a preference for long-unburnt habitat. They are generally absent from recently burnt habitat, but may forage in areas burnt 1-2 years earlier (Benshemesh 1990). They may not breed in burnt areas for up to 17 years after fire (Tarr 1965; Cowley *et al.* 1969) and are most abundant in long-unburnt habitat. For example, Benshemesh (1990, 1992) reported that breeding densities were three times greater in sites with post-fire ages of \geq 40 years than in neighbouring sites with post-fire ages of 20-30 years. Similarly, Woinarski (1989a, b) detected more Malleefowl in habitat with post-fire ages of 60-80 years earlier than in habitat with post-fire ages of 40 years or less.

Malleefowl require abundant ground-litter with which to construct their nest mounds (Benshemesh 2005). They are susceptible to fragmentation of their habitat: they appear to disperse on foot and require connections between patches of suitable habitat to facilitate movement (Benshemesh 1992, 2000, 2005).

In addition to their association with the properties discussed above, Malleefowl are highly susceptible to predation, especially by foxes (Frith 1959; Booth 1987; Priddel & Wheeler 1994, 1996, 1999; Benshemesh 2005). This susceptibility suggests that Malleefowl may offer some potential as an indicator species to monitor predator densities. However, Frith (1962) reported that Malleefowl could remain abundant in areas with high number of foxes, which if true would present problems for using presence/absence or abundance data on this species to indirectly assess predator numbers.

Mallee Emu-wren Stipiturus mallee

Natural Ecosystem Group: Mallee.

Properties Indicated: Fire regime; habitat connectivity; condition of Triodia grasslands.

Justification: Mallee Emu-wrens have been recorded in habitats with a range of post-fire ages (Silveira 1993). They can infiltrate burnt habitats as early as three years after fire (Garnett & Crowley 2000; Clarke 2005) and have been recorded breeding in high densities five years after fire (Garnett 1993), but their densities do not peak until 8-10 years after fire. They may sustain peak densities until 30 years after fire, at which time numbers begin to decline. However, some individuals may persist in habitats with post-fire ages of up to 50 years (Carpenter & Matthew 1986; Silveira 1993; Garnett & Crowley 2000; Clarke 2005; Mustoe 2006).

Mallee Emu-wrens are resident or sedentary birds (Rowley & Russell 1997) that have a limited ability to disperse across unsuitable habitat. This makes them susceptible to habitat fragmentation (Baker-Gabb 2005). Their minimum requirements for habitat connectivity have not been quantified, but they are capable of colonising recovering habitat at least 6 km away after fire (Garnett & Crowley 2000).

Yellow-throated Miner Manorina flavigula

Natural Ecosystem Group: Mallee.

Properties Indicated: Open or fragmented mallee.

Justification: Yellow-throated Miners avoid woody communities with dense and continuous vegetation. In regards to mallee communities, Yellow-throated Miners are typically recorded in open mallee or small mallee remnants that lie adjacent to open areas (*e.g.* natural clearings, roads or farmland). Yellow-throated Miners readily infiltrate formerly unsuitable areas when dense and continuous vegetation is fragmented or otherwise degraded (Joseph 1986; Emison *et al.* 1987; Starks 1987; Storr 1987; McLaughlin 1990; Franklin 1996).

Chestnut Quail-thrush Cinclosoma castanotum

Natural Ecosystem Group: Mallee.

Properties Indicated: Low shrubby undergrowth; well-developed leaf-litter; fire history.

Justification: Chestnut Quail-thrushes inhabit three main habitat types: (1) low shrubby undergrowth in woodlands dominated by mallee eucalypts or other tree species, (2) shrublands dominated by *Acacia* or *Melaleuca* and (3) heathlands (*e.g.* Ford 1970, 1971; Greenslade *et al.* 1986; Woinarski 1989b; Matthew & Carpenter 1990; Baxter & Paton 1998; Luck *et al.* 1999; Higgins & Peter 2002). They inhabit vegetation with post-fire ages ranging from 4-40 years, but are most abundant in more recently burnt habitats (http://www.threatenedspecies.environment.nsw.gov.au/tsprofile/profile.aspx?id=10168).

Red-kneed Dotterel Erythrogonys cinctus

Natural Ecosystem Group: Inland Waters and Wetlands.

Properties Indicated: Wetlands with fringing emergent vegetation and exposed sediment margins.

Justification: Red-kneed Dotterels frequent ephemeral or permanent wetlands with shallow water, emergent or fringing vegetation and margins of exposed mud or clay or sand. They are most common at freshwater and brackish wetlands, tending to avoid wetlands with more saline waters (e.g. McGill 1944; Hobbs 1961; Maclean 1977; Fjeldså 1985; Schulz 1986; Emison *et al.* 1987; Morris *et al.* 1990; Johnstone & Storr 1998; Lyons *et al.* 2007). Wetlands favoured by this species are often also suitable for a host of threatened migratory and resident shorebirds such as the Australian Painted Snipe *Rostratula australis*.

Red-capped Plover Charadrius ruficapillus

Natural Ecosystem Group: Coastal, Inland Waters and Wetlands.

Properties Indicated: Anthropogenic disturbance; predators.

Justification: Red-capped Plovers are sensitive to disturbance when nesting. Adult plovers flush from their nests when approached by humans or domestic animals, leaving the unprotected eggs or chicks vulnerable to predation (Lane 1987). Nests may be destroyed by humans or domestic animals and raided by predators such as foxes and gulls (Hobbs 1972; Lane 1987; Marchant & Higgins 1993). To most accurately determine the level of human disturbance or the presence or density of predators, monitoring should target breeding records (including rates of breeding success and causes of breeding failure) rather than simply presence/absence or abundance data.

Ground Parrot Pezoporus wallicus

Natural Ecosystem Group: Heathland.

Properties Indicated: Anthropogenic disturbance; fire regime.

Justification: In Victoria, Ground Parrots inhabit comparatively undisturbed, closed coastal graminoid or shrub heathlands and dense sedgelands (Isles & Menkhorst 1975; Meredith & Isles 1980; Meredith 1984; Meredith *et al.* 1984; Meredith & Jaremovic). They are adversely affected by degradation and fragmentation of their habitat (Meredith 1984; Baker 1997). For example, one area of formerly suitable habitat at Wilson's Promontory was abandoned after it was grazed and trampled by cattle (Cooper 1975).

Fire frequency is an important determinant of the suitability of heathland habitats for Ground Parrots. Fire eliminates vegetation that provides cover and food for Ground Parrots; frequent fires may prevent vegetation from recovering to a level suitable for Ground Parrots whereas infrequent fires may permit dense stands of tall woody shrubs to develop and replace low herbaceous vegetation favoured by Ground Parrots (Baker 1997). Ground Parrots may use recently burnt heathland vegetation that lies immediately adjacent to an established population in suitable habitat (McFarland 1989; Burbidge *et al.* 2007), but where such a situation does not exist, burnt heathlands may remain unsuitable for up to four years after fire (Garnett & Crowley 2000). Densities of Ground Parrots in heathlands peak 4-15 years after fire. Densities may decline to zero in some heathlands that have remained unburnt for 15-25 years (Meredith & Isles 1980; Jordan 1987a; Wall 1989; McFarland 1991; Bryant 1991), but

individuals may persist at lower densities in other heathlands that have remained unburnt for up to 90 years (Watkins 1985; Bryant 1991, 1994; Burbidge *et al.* 2007).

Southern Emu-wren Stipiturus malachurus

Natural Ecosystem Group: Heathland.

Properties Indicated: Comparatively undisturbed habitat; habitat connectivity; fire regime.

Justification: Southern Emu-wrens inhabit dense low vegetation in heathlands and other vegetation communities (Emison *et al.* 1987; Higgins *et al.* 2001; Wilson & Paton 2004; Maguire 2006). They are adversely affected by multiple forms of anthropogenic disturbance including clearing, fragmentation, drainage or salinasation of wetlands, grazing and altered fire regimes (Garnett 1993; Mount Lofty Ranges Southern Emu-wren Recovery Team 1998; Pickett 2002). They require corridors of dense vegetation to move between patches of suitable habitat (Pickett 2002).

Southern Emu-wrens are absent from recently burnt areas (Jordan 1987b; McFarland 1988). They may colonise regenerating heathland within one year and attain high densities within 2-3 years after fire (Jordan 1987b), although in heathland-woodland and swampland communities they may not return for more than three and more than four years after fire respectively (Reilly 1991). In Cooloola Park National in Queensland, Southern Emu-wrens occur in heathlands with post-fire ages of 1.5-9.0 years, but are present in greatest densities in heathlands with post-fire ages of 6-8 years (McFarland 1988, 1994). The results of this study suggest that the species is best suited by a regime of mosaic burns at intervals of 7-8 years (McFarland 1988).

Banded Lapwing Vanellus tricolor

Natural Ecosystem Group: Grasslands.

Properties Indicated: Open short grasslands; predator densities; .

Justification: Banded Lapwings occur in open, short native grasslands on treeless plains and in other endemic and modified open habitats with short ground-cover vegetation (Emison et al. 1987; Marchant & Higgins 1993; Johnstone & Storr 1998). As a ground-foraging and ground-nesting species, Banded Lapwings are vulnerable to terrestrial predators; monitoring of breeding success and causes of breeding failure may therefore provide some indication of local predator densities. Short, open native grasslands inhabited by Banded Lapwings are also the preferred habitat of the Plains-wanderer (Emison et al. 1987; Baker-Gabb 1990), which is listed as a threatened species under both federal government and Victorian state government legislation. Banded Lapwings and Plains-wanderers are considered to co-occur in some areas of Victoria, which suggests that the presence of Banded Lapwings on open, short native grasslands may be a useful indicator of Plains-wanderer habitat.

Sooty Owl Tyto tenebricosa

Natural Ecosystem Group: Wet Forest and Rainforest.

Properties Indicated: Tall (old-growth) eucalypts (including hollow-bearing trees); arboreal hollow-dependent mammals; large areas of continuous forest.

Justification: Sooty Owls inhabit large and intact areas of tall montane eucalypt forest. They are most abundant in old-growth forests, but they often also occur in younger forests that are located near old-growth forest stands or that contain patches of old trees and high densities

of stags (which provide habitat for hollow-dependent prey). Sooty Owls almost always nest and often roost in large hollows in tall trees.

A high proportion of their diet consists of arboreal, hollow-dependent mammals (Schodde & Mason 1980; Smith 1984a; Beruldsen 1986; Loyn *et al.* 1986; Emison *et al.* 1987; Hollands 1991; Milledge *et al.* 1991, 1993; Lundie-Jenkins 1993; Debus 1994; Holmes 1994; Kavanagh 1997; Kavanagh & Jackson 1997; Higgins 1999; Bilney *et al.* 2007). Sooty Owls have been proposed as an indicator species to guide the ecologically sustainable management of timber production forests in Australia (Milledge *et al.* 1991; Kavanagh *et al.* 2004).

Superb Lyrebird Menura novaehollandiae

Natural Ecosystem Group: Wet Forest and Rainforest.

Properties Indicated: Open ground cover; leaf-litter.

Justification: Superb Lyrebirds require an open forest floor with overlying leaf-litter for foraging (Ashton & Bassett 1997; Morgan et al. 1995; Higgins et al. 2001). As a terrestrial species, Superb Lyrebirds are vulnerable to terrestrial predators such as foxes, cats and dogs (Emison et al. 1987; Reilly 1988; Smith 1994, 1995). For example, one observer recorded 42 lyrebird tail-plumes in a single fox lair (Leach 1929) and predation by foxes was a primary factor in a decline in Superb Lyrebird numbers in Dandenong Ranges National Park (Smith 1994).

Red-browed Treecreeper Climacteris erythrops

Natural Ecosystem Group: Wet Forest and Rainforest.

Properties Indicated: Mature forest and rainforest; hollow-bearing (*i.e.* mature) trees.

Justification: Red-browed Treecreepers prefer mature forest but also occur less frequently in younger, regenerating stands (Milledge & Recher 1985; Loyn 1985; Smith 1984b, 1985a, b; Emison et al. 1987). They typically nest and roost in hollows of small to large, living or dead trees (Noske 1982a, 1985; Higgins et al. 2001). Red-browed Treecreepers were identified by Kavanagh et al. (2004) and Kavanagh & Stanton (2005) as candidate indicator species to monitor the impacts of logging in production forests because they depend on large, mature trees for nesting, foraging and roosting.

Rose Robin Petroica rosea

Natural Ecosystem Group: Wet Forest and Rainforest.

Properties Indicated: Tall mid-storey or understorey vegetation.

Justification: Rose Robins mainly occur in eucalypt forests and woodlands that contain a sparse to dense mid-storey or understorey of tall shrubs and low trees, especially species of Acacia (Loyn 1985, 1993; Emison et al. 1987; Gosper 1992; Slater 1995; Mac Nally 1997). They are occasionally recorded in coastal scrub (Bedggood 1980), mangroves (NSW Bird Report 1998) and modified habitats (e.g. residential gardens), the latter typically when on migration (Emison et al. 1987; Whiteside 1987; Templeton 1992).

Olive Whistler Pachycephala olivacea

Natural Ecosystem Group: Wet Forest and Rainforest.

Properties Indicated: Dense understorey vegetation.

Justification: Olive Whistlers usually occur in tall, dense undergrowth in a variety of wooded habitats (Loyn 1985; Emison et al. 1987; Higgins & Peter 2002).

Azure Kingfisher Ceyx azureus

Natural Ecosystem Group: Inland Waters and Wetlands, Wet Forest and Rainforest.

Properties Indicated: Overhanging branches; snags; earthen banks; conditions capable of sustaining aquatic prey species.

Justification: Azure Kingfishers inhabit well-vegetated wetlands with overhanging branches or tree-roots, or snags, logs or debris, that provide perches for loafing and foraging (Marshall 1931; Forshaw & Cooper 1983; Green 1995; Curl 1998 in Higgins 1999; Johnstone & Storr 1998). They are generally absent from wetlands that lack fringing vegetation (Barnard & Barnard 1925; Boekel 1980). They usually nest in burrows excavated in earthen banks on the margins of waterways or wetlands, or in soil clumped around roots of fallen trees (Marshall 1931; Bedggood 1973; Forshaw & Cooper 1983; Green 1995; Johnstone & Storr 1998; Curl 1998 in Higgins 1999). They sometimes nest in artificial banks (Curl 1998 in Higgins 1999).

Azure Kingfishers feed on aquatic vertebrates and invertebrates (Johnstone & Storr 1998; Higgins 1999). Their presence is therefore likely to indicate the existence of hydrological conditions capable of sustaining prey species.

The Azure Kingfisher was nominated by Kavanagh *et al.* (2004) as a potential target indicator species to monitor the impacts of logging in production forests.

Striated Fieldwren Calamanthus fuliginosus

Natural Ecosystem Group: Coastal.

Properties Indicated: Low, dense vegetation.

Justification: Striated Fieldwrens inhabit low, dense vegetation. They mainly occur in heathlands, sedgelands, samphire shrublands, saltmarsh and grassy swamplands, but are occasionally recorded in forests with dense understorey vegetation, grasslands, lightly-grazed pastures adjacent to more typical habitat, rank pastures and softwood plantations (Ridpath & Moreau 1966; Ratkowsky & Ratkowsky 1977; Thomas 1979; Emison et al. 1987; Gosper & Baker 1997; Higgins & Peter 2002). Although Striated Fieldwrens sometimes occur in disturbed habitats, they are adversely affected and/or threatened by degradation of their natural habitats through clearing, overgrazing and invasion by weeds (Napier 1969; Burbidge 1985; http://threatenedspecies.environment.nsw.gov.au/tsprofile/profile.aspx?id=10126).

Eastern Curlew Numenius madagascariensis

Natural Ecosystem Group: Coastal.

Properties Indicated: Anthropogenic disturbance of intertidal flats.

Justification: Eastern Curlews are sensitive to some forms of human disturbance (Close & Newman 1984; Peter 1990; Reid & Park 2003). For example, at Corner Inlet in Victoria,

Eastern Curlews took flight when approached by a single person on foot to a distance of 50-250 m, but a single bird failed to flush when approached by a boat to a distance of 20 m (Peter 1990). Despite this sensitivity, Eastern Curlews may not be a suitable indicator species for human disturbance, at least if analysis is based on presence/absence or abundance data, as Finn et al. (2007) found that the local level of human disturbance and the distance to urban development were both poor predictors of Eastern Curlew densities on intertidal flats at Moreton Bay in Queensland.

Australian Pied Oystercatcher Haematopus longirostris

Natural Ecosystem Group: Coastal.

Properties Indicated: Anthropogenic disturbance of sandy beaches.

Justification: Pied Oystercatchers are exposed to and adversely affected by disturbance from humans or associated recreational activities, e.g. nests, eggs or young may be destroyed by walkers, horses, domestic dogs or off-road vehicles (Hewish 1990; Newman 1990; McFarland 1993; Wilson 1994). The presence/absence or abundance of Pied Oystercatchers alone may not accurately reflect the existence or intensity of human disturbance. For example, the density of humans and 4WD vehicles appeared to have no significant influence on the distribution of Pied Oystercatchers along ocean beaches in northern New South Wales (Owner & Rohweder 2003). Instead, data analysis would need to focus on breeding records, including rates of breeding success and causes of breeding failure.

Hooded Plover Thinornis rubicollis

Natural Ecosystem Group: Coastal.

Properties Indicated: Anthropogenic disturbance of high-energy sandy beaches.

Justification: Hooded Plovers are highly vulnerable to anthropogenic disturbance (Dowling & Weston 1999; Weston 2003; Weston & Elgar 2005, 2007). There are several mechanisms by which anthropogenic disturbance may impact on Hooded Plovers:

- Adults may be flushed from their nests by humans or domestic dogs, leaving unprotected eggs or chicks vulnerable to predation or to thermal or energetic stress (Schulz 1992; Retallick & Bolitho 1993; Weston 2000).
- Nests or young may be crushed by humans, vehicles, domestic dogs, privately-owned horses and horses and camels used in commercial tourism operations (Buick & Paton 1989; Schulz 1992; Weston 1993, 2003; Dowling & Weston 1999).
- Adults may be killed by collisions with vehicles (Rose 2000; Weston & Morrow 2000).
- Nests may be interfered with by humans, e.g. mischievous or well-meaning collection of eggs or chicks (Schulz 1992; Weston 2000, 2003).
- Eggs, young or adults may be preyed upon by domestic dogs (Retallick & Bolitho 1993; Hanisch 1998; Weston 1998, 2003; Weston & Morrow 2000).
- Scavengers may be attracted to breeding sites by discarded litter (Schulz & Bamford 1987).

It is likely that monitoring breeding success and causes of breeding failure would provide a more accurate surrogate assessment of human disturbance than analysing presence/absence or abundance data alone. Monitoring of breeding success is regularly undertaken in various coastal sites in Victoria by Birds Australia and volunteers.

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APPENDIX 1

Table A1.1 Summary of identified indicator or focal bird species in Australia from published studies.

Location	Habitat	Indicator Species	Indicated	Source
South-western Western Australia.	Jarrah <i>Eucalyptus marginata</i> , Wandoo <i>E. wandoo</i> and Karri <i>E. diversicolor</i> forests.	Live, hollow-bearing trees: Red-tailed Black-Cockatoo <i>Calyptorhynchus banksii</i> and Rufous Treecreeper <i>Climacteris rufa</i> . Riparian forest: Splendid Fairy-wren <i>Malurus splendens</i> , Red-winged Fairy-wren <i>Malurus elegans</i> , White-breasted Robin <i>Eopsaltria georgiana</i> and Red-eared Firetail <i>Stagonopleura oculata</i> . Older regrowth and old growth forest: White-browed Babbler <i>Pomatostomus superciliosus</i> , Black-faced Cuckoo-shrike <i>Coracina novaehollandiae</i> and Restless Flycatcher <i>Myiagra inquieta</i> . Fire history: Golden Whistler <i>Pachycephala pectoralis</i> .	Live, hollow-bearing trees; riparian forest; older regrowth and old growth forest; fire history.	Abbott (1999) ¹
South-western Western Australia.	Jarrah, Wandoo and Karri forests.	Red-tailed Black-Cockatoo, Baudin's Black-Cockatoo Calyptorhynchus baudinii, Western Rosella Platycercus icterotis, Red-capped Parrot Purpureicephalus spurius, Sacred Kingfisher Todiramphus sanctus and Rufous Treecreeper.	Live, hollow-bearing trees.	Abbott & Whitford (2002) ²
South-western Western Australia.	Heathland/shrubland/mallee and woodland.	Heathland/shrubland/mallee patch size: Shy Heathwren <i>Hylacola cauta</i> (Redthroat <i>Pyrrholaemus brunneus</i> and Crested Bellbird <i>Oreoica gutturalis</i>). Heathland/shrubland/mallee patch isolation: Southern Scrub-robin <i>Drymodes brunneopygia</i> (Redthroat). Woodland patch size: Varied Sittella <i>Daphoenositta chrysoptera</i> and Dusky Woodswallow <i>Artamus cyanopterus</i> (Varied Sittella). Woodland patch isolation: Varied Sittella (Varied Sittella). Woodland condition: Varied Sittella (Varied Sittella). Remnant size: Jacky Winter <i>Microeca fascinans</i> (White-eared Honeyeater <i>Lichenostomus leucotis</i>). Remnant condition: White-eared Honeyeater (White-eared Honeyeater).	Heathland/shrubland/mallee patch size; heathland/shrubland/mallee patch isolation; woodland patch size; woodland patch isolation; woodland condition; remnant size; remnant condition.	Brooker (2002) ³
South-western Western Australia.	Woodland and shrubland.	Woodland: Rufous Treecreeper. Shrubland: Shy Heathwren. Generalist: Grey Currawong Strepera versicolor.	Habitat fragmentation.	Brooker & Lefroy (2004) ⁴
South-western Western Australia.	Woodland, shrubland and heathland.	Woodland patch area: Varied Sitella. Shrubland patch area: Spiny-cheeked Honeyeater, Western Yellow Robin <i>Eopsaltria griseogularis</i> and Southern Scrubrobin. Heathland patch area: Redthroat.	Woodland patch area; shrubland patch area; heathland patch area.	Frost <i>et al.</i> (1999) ⁵
South-western Western Australia.	Heathland/shrubland/mallee and woodland.	Heathland/shrubland/mallee patch size: Southern Scrub-robin. Heathland/shrubland/mallee patch isolation: Southern Scrub-robin. Woodland patch size: Red Wattlebird <i>Anthochaera carunculata</i> . Woodland patch isolation: Brown- headed Honeyeater <i>Melithreptus brevirostris</i> . Remnant area: Grey Butcherbird <i>Cracticus torquatus</i> . Remnant condition: Western Yellow Robin.	Heathland/shrubland/mallee patch size; heathland/shrubland/mallee patch isolation; woodland patch size; woodland patch isolation; remnant area; remnant condition.	Huggett <i>et al.</i> (2004)

Location	Habitat	Indicator Species	Indicated	Source
South-western Western Australia.	Woodland, shrubland and heathland.	Woodland patch area: White-eared Honeyeater, Golden Whistler, Restless Flycatcher and Jacky Winter. Shrubland patch area: Western Thornbill Acanthiza inornata and Southern Scrub-robin. Heathland patch area: White-browed Scrubwren, Shy Heathwren, Western Thornbill, Tawny-crowned Honeyeater Glyciphila melanops and White-cheeked Honeyeater Phylidonyris nigra.	Woodland patch area; shrubland patch area; heathland patch area.	Lambeck (1998)
South-western Western Australia.	Woodlands, shrublands/mallee, shrublands/heathlands.	Habitat loss in woodland: Jacky Winter and Varied Sittella. Habitat loss in shrubland/mallee: Shy Heathwren. Habitat loss in shrubland/heathland: Rufous Fieldwren <i>Calamanthus campestris</i> . Fragmentation: Shy Heathwren, Rufous Fieldwren and Western Yellow Robin.	Habitat loss; habitat fragmentation. [Lambeck cites honeyeaters as being resource-limited, but am not sure that this is a valuable property for an indicator as to be used by Parks Victoria?]	Lambeck (1999)
South-western Western Australia.	Tall open Karri forests.	Live, hollow-bearing trees: Red-tailed Black-Cockatoo and Baudin's Black- Cockatoo. Old regrowth and old growth forest: Red-tailed Black-Cockatoo, Baudin's Black-Cockatoo, Red-capped Parrot, Fan-tailed Cuckoo <i>Cacomantis flabelliformis</i> , Rufous Treecreeper, White-browed Babbler, Dusky Woodswallow <i>Artamus</i> <i>cyanopterus</i> and Tree Martin.	Live, hollow-bearing trees; older regrowth and old growth forest.	Williams <i>et al.</i> (2001) ⁹
Arid rangelands of Western Australia, Northern Territory, South Australia, Queensland and New South Wales.	Chenopod and acacia shrublands.	Birds (individual species not identified).	Grazing impacts.	Landsberg <i>et al.</i> (1999) ⁷
Top End of the Northern Territory.	Coastal monsoon rainforest and associated coastal savannas.	Orange-footed Scrubfowl Megapodius reinwardt (nest mounds).	Habitat change.	Bowman <i>et al.</i> (1994)
South-eastern Queensland and north-eastern New South Wales.	Woodlands, wetlands.	Woodlands: Brown Treecreeper <i>Climacteris picumnus</i> , Variegated Fairy-wren, White-winged Fairy-wren <i>Malurus leucopterus</i> , Speckled Warbler, Spiny-cheeked Honeyeater, Striped Honeyeater, Grey-crowned Babbler, Golden Whistler, Hooded Robin <i>Melanodryas cucullata</i> , Eastern Yellow Robin and Diamond Firetail. Wetlands: Purple Swamphen <i>Porphyrio porphyrio</i> .	Biodiversity condition.	Ford & Thompson (2006)
South-east Queensland.	Softwood scrub remnants/farmland and open eucalypt forest.	Habitat change in softwood remnants/farmland: Crested Pigeon Ocyphaps lophotes, Peaceful Dove, Pale-headed Rosella Platycercus adscitus, Variegated Fairy-wren, White-browed Scrubwren, Speckled Warbler, Varied Triller Lalage leucomela, Australasian Figbird Sphecotheres vieilloti, Rufous Whistler Pachycephala rufiventris, Willie Wagtail, Eastern Yellow Robin, Mistletoebird and Zebra Finch Taeniopygia guttata. Habitat change in open eucalypt forest: Peaceful Dove, Buff-rumped Thornbill, Noisy Miner, Rufous Whistler, Olive-backed Oriole, Grey Butcherbird, Pied Butcherbird Cracticus nigrogularis, Pied Currawong, Eastern Yellow Robin and Mistletoebird.	Habitat change.	Leach (1996) ⁸

Location	Habitat	Indicator Species	Indicated	Source
Queensland, New South Wales, Victoria and Tasmania.	Forests.	Wonga Pigeon Leucosarcia picata, White-throated Nightjar Eurostopodus mystacalis, Bush Stone-curlew Burhinus grallarius, Red-tailed Black-Cockatoo, Glossy Black-Cockatoo Calyptorhynchus lathami, Yellow-tailed Black-Cockatoo C. funereus, Gang-gang Cockatoo Callocephalon fimbriatum, Musk Lorikeet Glossopsitta concinna, Little Lorikeet G. pusilla, Australian King-Parrot Alisterus scapularis, Superb Parrot Polytelis swainsonii, Green Rosella Platycercus caledonicus, Crimson Rosella, Swift Parrot Lathamus discolor, Powerful Owl, Barking Owl Ninox connivens, Southern Boobook N. novaeseelandiae, Sooty Owl, Masked Owl Tyto novaehollandiae, Azure Kingfisher Ceyx azureus, White-throated Treecreeper, Red-browed Treecreeper Climacteris erythrops, Brown Treecreeper, Speckled Warbler, Buff-rumped Thornbill Acanthiza reguloides, Striated Pardalote, Regent Honeyeater Anthochaera phrygia, Strong-billed Honeyeater Melithreptus validirostris, Brown-headed Honeyeater, White-naped Honeyeater M. lunatus, Black-headed Honeyeater M. affinis, Painted Honeyeater Grantiella picta, Grey- crowned Babbler, White-browed Babbler, Spotted Quail-thrush Cinclosoma punctatum, Varied Sittella Daphoenositta chrysoptera, Cicadabird Coracina tenuirostris, Crested Shrike-tit, Olive-backed Oriole Oriolus sagittatus, Leaden Flycatcher Myiagra rubecula, Satin Flycatcher M. cyanoleuca, Scarlet Robin Petroica boodang, Red-capped Robin Petroica goodenovii, Hooded Robin, Tree Martin Hirundo nigricans and Mistletoebird.	Logging impacts.	Kavanagh <i>et al.</i> (2004)
North-eastern New South Wales.	Assorted forests.	Grey Goshawk Accipiter novaehollandiae, Collared Sparrowhawk A. cirrhocephalus, Little Lorikeet, Crimson Rosella, Eastern Rosella, Albert's Lyrebird Menura alberti, Red-browed Treecreeper, White-throated Gergyone Gerygone olivacea, Buff- rumped Thornbill, Little Wattlebird Anthochaera chrysoptera, Yellow-tufted Honeyeater Lichenostomus melanops, Pale-yellow Robin Tregellasia capito, Varied Sittella, Golden Whistler, Leaden Flycatcher, Satin Flycatcher, Willie Wagtail, Spangled Drongo Dicrurus bracteatus, Torresian Crow Corvus orru and Bassian Thrush Zoothera lunulata or Russet-tailed Thrush Z. heinei more abundant in unlogged forest; and Australian Brush-turkey Alectura lathami, Brown Goshawk Accipiter fasciatus, Brown Cuckoo-Dove Macropygia amboinensis, Wonga Pigeon, Rose-crowned Fruit-Dove Ptilinopus regina, Shining Bronze-Cuckoo Chrysococcyx lucidus, Channel-billed Cuckoo Scythrops novaehollandiae, White-browed Scrubwren, White-throated Needletail Hirundapus caudacutus, Brown Gerygone Gerygone mouki, Bell Miner Manorina melanophrys, Lewin's Honeyeater Meliphaga lewinii, Yellow-faced Honeyeater, White-throated Honeyeater Meliphaga lewinii, Yellow-faced Honeyeater, White-throated Honeyeater Meliphaga lewinii, Seastern Spinebill Acanthorhynchus tenuirostris, Scarlet Honeyeater Myzomela sanguinolenta, Eastern Whipbird Psophodes olivaceus, Little Shrike- thrush Colluricincla megarhyncha, Black-faced Monarch Monarcha melanopsis, Rufous Fantail Rhipidura rufifrons, Olive-backed Oriole, Figbird Sphecotheres viridis, Pied Currawong Strepera graculina, Paradise Riflebird Ptilonorhynchus violaceus more abundant in logged forest.	Logging impacts.	Kavanagh & Stanton (2005)
South-central New South Wales.	Woodland remnants.	Southern Whiteface <i>Aphelocephala leucopsis</i> , White-browed Babbler and Hooded Robin.	Remnant size; remnant isolation; habitat complexity.	Freudenberger & Stol (2002)

Location	Habitat	Indicator Species	Indicated	Source
South-central New South Wales.	Floodplain (riparian) eucalypt woodlands.	Grazing activity: Galah <i>Eolophus roseicapillus</i> , Sulphur-crested Cockatoo <i>Cacatua</i> <i>galerita</i> , Eastern Rosella <i>Platycercus eximius</i> , Australian Magpie Cracticus tibicen, Willie Wagtail <i>Rhipidura leucophrys</i> , Magpie-lark <i>Grallina cyanoleuca</i> , Rufous Songlark <i>Cincloramphus mathewsi</i> and Common Starling <i>Sturnus vulgaris</i> more abundant at heavily grazed sites; and Peaceful Dove <i>Geopelia striata</i> , Crimson Rosella <i>Platycercus elegans</i> , Red-rumped Parrot <i>Psephotus haematonotus</i> , Laughing Kookaburra <i>Dacelo novaeguineae</i> , White-throated Treecreeper <i>Cormobates leucophaea</i> , Brown Treecreeper, Superb Fairy-wren <i>Malurus cyaneus</i> , Variegated Fairy-wren, Weebill <i>Smicrornis brevirostris</i> , Striated Pardalote <i>Pardalotus striatus</i> , Noisy Friarbird <i>Philemon corniculatus</i> , Little Friarbird <i>P. citreogularis</i> , Grey Shrike-thrush <i>Colluricincla harmonica</i> , Black-faced Cuckoo- shrike, Grey Fantail <i>Rhipidura albiscapa</i> , Australian Raven <i>Corvus coronoides</i> , White-winged Chough <i>Corcorax melanorhamphos</i> and Mistletoebird <i>Dicaeum</i> <i>hirundinaceum</i> more abundant at least grazed sites. Clearing: Galah, Sulphur- crested Cockatoo, Eastern Rosella, Noisy Miner <i>Manorina melanocephala</i> , Australian Magpie, Willie Wagtail, Magpie-lark, Rufous Songlark and Common Starling more abundant at least forested sites; and Peaceful Dove, Crimson Rosella, Red-rumped Parrot, Laughing Kookaburra, White-throated Treecreeper, Brown Treecreeper, Superb Fairy-wren, Variegated Fairy-wren, White-browed Scrubwren <i>Sericornis frontalis</i> , Striated Pardalote, Noisy Friarbird, Little Friarbird, Black-faced Cuckoo-shrike, Crested Shrike-tit <i>Falcunculus frontatus</i> , Grey Shrike- thrush, Grey Fantail, Australian Raven, White-winged Chough and Mistletoebird more abundant at most forested sites.	Grazing activity; clearing.	Jansen & Robertson (2001) ⁶
South-eastern New South Wales.	Woodlands.	Eastern Yellow Robin.	Habitat loss; habitat complexity; habitat fragmentation.	Freudenberger (2001)
South-eastern New South Wales.	Assorted forests.	Powerful Owl Ninox strenua and Sooty Owl Tyto tenebricosa.	Old-growth forests with large, hollow-bearing trees.	Kavanagh (1991)
South-eastern Australia.	Open eucalypt forests.	Yellow-tufted Honeyeater Lichenostomus melanops and Bell Miner Manorina melanophrys.	Gullies and riparian areas of wet forest; biodiversity.	Neave <i>et al.</i> (1996)
Australian Capital Territory.	Eucalypt woodlands.	Habitat loss: Hooded Robin. Habitat fragmentation: Eastern Yellow Robin. Habitat complexity: Hooded Robin.	Habitat loss; habitat fragmentation; habitat complexity.	Watson <i>et al.</i> (2001)
Central Victorian highlands.	Mountain Ash <i>Eucalyptus regnans</i> forest.	Sooty Owl.	Old-growth forest with large, hollow-bearing trees.	Milledge <i>et al.</i> (1991)
South-east Tasmania.	Ocean.	Fairy Prion Pachyptila turtur.	Heavy metal pollution.	Brothers & Brown (1987)

¹ The absence of established populations of species typical of disturbed or sparsely-wooded habitats (*e.g.* Black-shouldered Kite *Elanus axillaris*, Nankeen Kestrel *Falco cenchroides*, Singing Honeyeater *Lichenostomus virescens*, White-fronted Chat *Epthianura albifrons*, White-winged Triller *Lalage sueurii*, Grey Butcherbird, Willie Wagtail, Magpie-lark, Red-capped Robin, Australasian Pipit *Anthus novaeseelandiae*) may also indicate habitat condition.

² Two mammal species, Brush-tailed Possum *Trichosurus vulpecula* and Brush-tailed Phascogale *Phascogale tapoatafa*, were also proposed as candidate indicators. Of the eight candidate indicators identified, Brush-tailed Possum was considered likely to provide the earliest indication of change.

³ The most sensitive species for each habitat variable is presented in the 'Candidate Indicator' column followed in brackets by the critical focal species that were chosen by Brooker to meet the aim of the study (*i.e.* to retain existing resident avifauna of the Gabbi Quoi Quoi region). Other focal (sensitive) species identified were Blue-breasted Fairy-wren *Malurus pulcherrimus*, Inland Thornbill *Acanthiza apicalis*, White-browed Babbler, Golden Whistler, Western Yellow Robin and Southern Scrub-robin (heathland/shrubland/mallee patch size); Blue-breasted Fairy-wren (heathland/shrubland/mallee patch isolation); Rufous Treecreeper, Weebill, Red Wattlebird, Brown-headed Honeyeater, Grey Currawong and Red-capped Robin (woodland patch size); Rufous Treecreeper and Red-capped Robin (woodland patch isolation); Brown-headed Honeyeater (woodland patch condition); Common Bronzewing *Phaps chalcoptera*, Chestnut-rumped Thornbill *Acanthiza uropygialis*, Brown Honeyeater *Lichmera indistincta*, Rufous Whistler *Pachycephala rufiventris*, Grey Shrike-thrush, Grey Butcherbird and Red-capped Robin (remnant size); and Common Bronzewing, Brown Honeyeater and Grey Shrike-thrush (remnant condition).

⁴ Other focal (sensitive) species identified were Brown-headed Honeyeater, Varied Sittella, Jacky Winter and Hooded Robin (woodland); Malleefowl, Blue-breasted Fairy-wren, White-browed Scrubwren, Redthroat, Golden Whistler, Crested Bellbird, Western Yellow Robin and Southern Scrub-robin (shrubland); and Bush Stone-curlew, Inland Thornbill and White-eared Honeyeater (generalist).

⁵ Other sensitive species identified were White-eared Honeyeater and Rufous Whistler (woodland patch area); and Shy Heathwren and Rufous Fieldwren (heathland patch area).

⁶ Brown Treecreeper and Superb-fairy Wren were considered the most suitable candidates for monitoring because they were found at most study sites and their abundance appeared to vary with grazing pressure.

⁷ Understorey plants were considered to be the most useful indicators of grazing impacts of the taxa examined (= understorey, overstorey and seed-bank plants; arthropods; reptiles, birds and small mammals). Birds were considered to be the most useful indicators amongst the faunal groups studied.

⁸ This study also recommended that non-migratory species with fluctuating population sizes (*e.g.* Bar-shouldered Dove *Geopelia humeralis* and Yellow Thornbill *Acanthiza nana*) be considered as potential indicators.

⁹ Other potential indicator species identified by this study were species usually absent from or vagrant in undisturbed forest (*e.g.* Laughing Kookaburra *Dacelo novaeguineae*, Splendid Fairy-wren, Brown Honeyeater and Australasian Pipit).

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